



Article Sustainable Italian Cities: The Added Value of Biomethane from Organic Waste

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Abstract: This work focuses on the profitability of biomethane plants and the environmental benefits obtained recovering the organic fraction of municipal solid waste in Italy. The economic model is based on the calculations of the net present value, considering multiple capacities of biomethane production (ranging from 50 to 500 m³/h) and alternative scenarios based on the variation in subsidies, the selling price of biomethane, and the net revenues from the treatment of organic waste. The environmental analysis quantifies the reduction in greenhouse gas emissions obtained by natural gas vehicles fueled by biomethane. The economic and environmental results encourage energy change that can be achieved by municipalities that support the transformation of natural resources into green fuels. Across 15 Italian municipalities, the potential biomethane production varies from 80.4 million m³/year to 102.8 million m³/year, with an overall net present value ranging from 135 to 187 million $CO_2eq/year$. Both the economic and environmental results demonstrate that biomethane is a renewable resource with added value for municipalities.

Keywords: biomethane; economic analysis; environmental analysis; organic waste; renewable energy; vehicle fuel

1. Introduction

A sustainable society requires members to take responsibility for managing waste in order to prevent adverse impacts on public health and the environment [1,2]. In 2016, an estimated 1.6 billion tons of carbon dioxide-equivalent (CO2eq) greenhouse gas (GHG) emissions were generated from solid waste management, representing approximately 5% of the global emissions [3]. The integration of renewable energy resources and sustainable development is fundamental for future generations [4,5]. There is a high relationship between the concepts of the green economy and the circular economy (CE) [6]. The CE model involves numerous sectors and aims at reducing both virgin material inputs and waste outputs [7]. CE can be viewed as a model that seeks to achieve sustainability through a special focus on economic and environmental dimensions [8]. Waste management strategies have the final aim of reducing waste generation and maximizing the separate collected phase [9]. Additionally, the transformation of waste into energy supports the development of a CE model by closing the loop-not only in terms of materials, but also in terms of energy [10]. For this reason, CE practices are relevant for the implementation of the Sustainable Development Goals [11]. The organic fraction of municipal solid waste (OFMSW) is recognized as a sustainable resource for renewable energy production, and anaerobic digestion (AD) is considered the best technological method for this process [12], as it reduces the generation of waste and maximizes the recycling and recovery of waste, thus minimizing the use of landfills [13].

The biogas-biomethane chain provides a double advantage, as it simultaneously produces energy and recovers residues, reducing the amount of material destined for the landfill [14,15] (Figure 1). Biogas is the product of AD from several feedstocks. Biomethane, in contrast, is produced by biogas upgrading, which is a renewable energy source (RES) suitable for electricity and heat production, injection into the gas grid, and for use as fuel for transportation [16]. The technical standards necessary for using biomethane (as an alternative to natural gas) are specific to the final destination [17]. Hydrogen is a clean fuel with no harmful emissions, and can be applied in fuel cells to generate electricity [18]. Biomethane supports the decarbonization of the transport system, involving both private and public actors [19,20]. The integration of several types of RESs can help to achieve sustainability [21], and provides opportunities for additional sources of income [22,23].

Europe demonstrates a great potential for biogas [24], with an estimated production from collectible manures of 27 billion m³ biogas (16.1 billion m³ biomethane) in the European Union (EU) alone [25]. Research has also estimated a production value of 18 billion m³ biomethane, considering a moderate technology penetration scenario and a large share of this potential used in the transport sector [26].

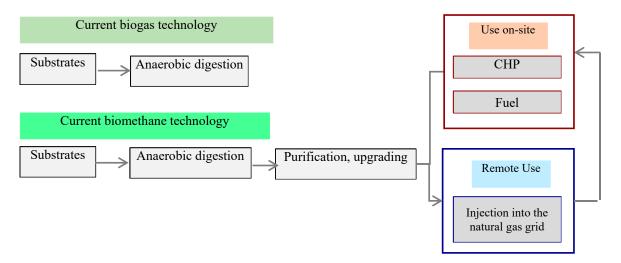


Figure 1. Map of the biogas-biomethane chain. Adapted from the literature [27].

It is expected that the global energy system will be decarbonized in order to meet the terms of the 2015 Paris Agreement. To this end, the European Commission has fixed targets and policy objectives for 2030, within the 2030 climate and energy framework, as follows: (i) at least 40% cuts in GHG emissions (from 1990 levels) and (ii) at least a 32% share of RES and a 32.5% improvement in energy efficiency [28]. The interconnection between renewable energy and energy efficiency was explored by the authors of [29]. A revised legislative proposal on waste was presented within the CE framework, recommending the following: (i) a common EU target for recycling of 65% of all municipal waste and (ii) a binding landfill cap of 10% municipal waste [30]. According to European legislation, the target shares of renewables in the transport sector (RES-T) will be fixed to 10% for all member states in 2020. The positive impact of biofuels on sustainable development was defined by the authors of [31]. In supporting this shift, however, government initiatives play a key role. Policies vary significantly in their tactics to promote the green energy transition in times of crisis, according to the specific aim pursued (i.e., profit, jobs, or EU targets) [32,33]. Sweden demonstrates best practice among member states in the development of green fuels [34], and some authors have focused their attention on the relationship between achieving regulatory compliance and improving biofuel sustainability [35]. In Sweden, public organizations make a significant contribution to the shift—some municipalities use environmentally friendly cars, Swedavia steers via a taxi queuing system, and Stockholm Public Transport procures biogas, matching the production of municipal wastewater treatment plants [19].

The development of CE models is verified by the recovery of some wastes that are perceived as resources and not only issues. Agricultural residues [36] and organic waste [37] are two such examples.

Furthermore, some researchers have demonstrated that, while OFMSW management is complex, OFMSW can be valorized through transformation into biomethane [38]. The co-digestion of multiple feedstocks is technically feasible, and the best method for enhancing biogas production is to combine the potential results of biomethane with continuous AD testing [39]. Energy crops present a significant methane potential, and they have been widely used in biogas plants; however, their use may determine a change in land use, which would penalize agriculture [40]. The use of residual biomass achieves a better environmental performance than the use of dedicated biomass [41]. Within this context, Italy represents an interesting case study, because of its developed biogas market and its undeveloped biomethane market [27,42]. Years ago, the Italian Biogas Consortium elaborated a roadmap for the production of approximately 8 billion Nm³/year biomethane [43], and the Italian Biomethane Decree (signed 2 March 2018) aims at promoting the production of biomethane in order to increase the share of green fuels in the Italian transportation sector [44].

The novelty of this work is based on the following assumptions: (i) biomethane is a promising renewable resource that may support decarbonization in the transport sector; (ii) single cities can promote CE models to increase the sustainability of the energy sector; (iii) quantitative analyses may provide valuable information to investors and policymakers; (iv) Italy is a paradoxical case study, as it has all of the available input for a biomethane market, but the development of this market has not yet occurred; and (v) biomethane supports the interaction between renewable energy and waste management. Focusing on 15 Italian municipalities, the research uses both economic and environmental analyses to demonstrate the added value of biomethane from organic waste. The paper presents a comparative analysis of four scenarios, including the current one. Starting with the amount of OFMSW produced in each geographic entity, a strategic analysis is introduced to define the optimal plant size in order to manage these flows. The production decision is necessarily long-term, because the lifetime of a biomethane plant is 20 years [27,45], and the interaction between renewable energy and waste management seeks to achieve a CE model. Subsequently, an economic analysis based on discounted cash flow (DCF) is proposed, and the index of the reference is the net present value (NPV). Baseline and alternative case studies are constructed on the basis of the defined critical variables. The value of the subsidy is fixed by decree through the Certificates of Emission of Biofuel in Consumption (CIC) for the first 10 years, but their value is not known for the last 10 years. This variable, as well as the selling price of biomethane and the net revenues linked to the treatment of municipal waste, significantly influence the discounted cash inflows; thus, all of these variables are considered critical variables.

In addition, a joint implementation model between a biomethane producer and a methane distributor is also considered. The analysis initially elaborates the "As Is" scenario, and subsequently explores three "To Be" scenarios. Regarding the environmental analysis, the reduction of GHG emissions is estimated on the basis of the literature data on the use of green gas (BIO-CNG) as an alternative to compressed natural gas (CNG). The calculation of the reduction in GHG emissions through NGVs (RGHG_{NGV}) also depends on the productivity of biomethane.

2. Literature Review

Generally, the biodegradable fraction of MSW varies from 30% to 65% [46], and OFMSW can be recovered from AD because of its high volatile solids [47]. AD supports the valorization of organic waste by generating biogas; for this reason, it is a valid option for the environmental management of MSW [48]. Using energy derived from biogas (as opposed to energy generated from fossil fuels) reduces CO₂ emissions by 200 to 300 kgCO₂/ton of biowaste [49]. Using energy derived from biomethane (BIO-CNG) follows the same approach, and its positive impact can be defined as follows:

- Reduced GHG emissions in vehicle power train systems of 119 gCO₂eq/km, 151 gCO₂eq/km, and 159 gCO₂eq/km, relative to compressed natural gas (CNG), diesel, and petrol, respectively [50].
- GHG emissions from biomethane varying from -36 to 10 gCO₂/MJ, all of which are lower than the 27 gCO₂/MJ produced by a conventional biogas plant. This reduction is even more significant, considering the value of natural gas in Europe is equal to 72 gCO₂/MJ [51].

• GHG emissions from biomethane production ranging from 45 to 68 gCO₂eq/kWh, all of which are lower than the 124 gCO₂eq/kWh, which is linked to the production of natural gas [52].

The economic perspective can be evaluated from several points of view (Table 1). Some studies have measured the cost of biomethane production, which varies from 0.54 to $0.73 \notin m^3$ if it is injected into the gas grid or used as transportation fuel, respectively [17], and from 0.71 to $0.75 \notin m^3$ if it is used for feedstock composition [53]. In particular, some authors have defined the cost of biogas production as mainly a function of the type of feedstock used (ranging from 0.38 to $0.56 \notin m^3$) [54]. The substrate and size also influence the cost of biomethane, which is estimated to range from $0.5 \text{ US/m^3} (4.7 cent \$US/kWh) to $1.5 \text{ US/m^3} (15 cent \$US/kWh) [55].

Other research has compared the costs of biomethane and natural gas. Currently, subsidies are needed in order to promote the development of the biomethane market [56]. The carbon price plays a key role in the decarbonization of energy systems, and competitiveness can be reached when the price falls within the following values: $70 \notin$ /ton of CO₂ [57] to $99 \notin$ /ton of CO₂ [58]. However, these values are significantly lower than the ones registered in another work, which found the costs to range from 180 to 260 \notin /ton of CO₂ [59]. Consequently, the analysis of biomethane plants is extremely complex, because multiple variables must be considered.

Some researchers have framed the difference between biogas and biomethane plants in terms of profitability. In fact, initially, several policymakers adopted subsidies only for the production of electricity from AD. Biomethane can be more convenient when there is both a high retail price and a low biogas upgrading cost [60]. Within existing biogas plants, the profitability of biomethane is typically reached when a centralized upgrading system is constructed. However, this is only valid when the biogas plants are in relatively close proximity to each other [45].

Indicator	Value	Reference
Biomethane cost production	0.54–0.73 €/m ³	[17]
Biomethane cost production	0.71–0.75 €/m ³	[53]
Biogas cost production	0.38–0.56 €/m ³	[54]
Biomethane cost production	0.5–1.5 \$US/m ³	[55]
Carbon price (subsidies)	70 €/ton of CO ₂	[57]
Carbon price (subsidies)	99 €/ton of CO ₂	[58]
Carbon price (subsidies)	180–260 €/ton of CO ₂	[59]
Net present value	(-585)-5667 K\$US	[61]
Net present value	0.49–132.7 M€	[62]
Net present value	616 K€	[63]
Net present value	(−341)−1174 K€	[64]
Net present value	1498–3200 K€	[65]
Discounted payback time	3–5 years	[65]
Net present value	(-2412)-2868 K€	[27]
Discounted payback time	<1 year	[27]

Table 1. Literature review—economic analysis of biomethane plants. USD—United States Dollar.

Finally, the profitability of biomethane plants has been investigated in several other works. In one study, the co-digestion of food waste and sewage sludge to produce biomethane was found to be financially feasible (NPV equal to 5667 K\$US), while that of food waste and pig slurry was found to be unprofitable (NPV equal to -585 K\$US) when no subsidies were provided [61]. Another study confirmed this result—NPV was found to vary from 89.8 M€ to 132.7 M€ when the organic waste had a consistent weight in the mix of waste recovered. However, if only pig slurry, manure residues, cattle slurry, milk processing waste, and slaughterhouse waste were used, the economic performance ranged from 0.49 M€ to 19.1 M€. The profitability was determined by the application of subsidies [62].

Plant size, in addition to subsidies and substrates, has also been shown to significantly influence financial feasibility. The minimum plant size in which profitability has been verified is 150 m³/h, using

the OFMSW as a substrate. At this size, the NPV has been calculated as 616 K€ when biomethane is injected into the gas grid [63]. Considering a CIC value of approximately $375 \notin$ (about $0.305 \notin$ /m³) and a selling price of biomethane equal to $0.1722 \notin$ /m³, the NPV is 165 K€. However, when the selling prices of biomethane are $0.1384 \notin$ /m³ and $0.2397 \notin$ /m³, respectively, the NPV is instead equal to $-341 \quad$ K€ and 1174 K€, respectively. In contrast, if a mixed substrate (energy crops and manure residues) is used, the minimum plant size is 500 m³/h, but only when vehicle fuel is the final destination of the biomethane [64]. Italy has often been analyzed as a case study in the literature, because of its great potential. Some authors have measured the performance of the upgrading technology used, with the NPV shown to vary from 1498 to 3200 K€, and both the payback time (PBT) and discounted payback time (DPBT) varying from 3 to 5 years [65]. Other authors have evaluated a 250 m³/h plant as being profitable when organic waste is used, but not when by-products or mixed (by-products and energy crops) substrates are used. Here, the NPV has been shown to vary from -2412 to 2868 K€, with a DPBT of potentially less than 1 year [27].

3. Materials and Methods

The realization of a new plant requires environmental and economic goals to be achieved. This section begins by describing the current production of MSW in Italy (Section 3.1). It then presents the strategic analysis, in which a biomethane plant size is proposed for each municipality. This calculation depends on the quantity of feedstock that must be treated, and, consequently, alternative scenarios are considered (Section 3.2). While the first decree towards the development of biomethane was published in December 2013, a new policy framework was recently launched for the Italian territory; these policies are discussed in Section 3.3. The profitability analysis of biomethane plants is conducted according to the economic model proposed in Section 3.4. Finally, Section 3.5 presents the input assumptions defined in this work, and Section 3.6 concerns the environmental analysis.

3.1. Municipal Solid Waste

As part of its Circular Economy Action Plan, the European Commission adopted a package of legislative proposals on waste in December 2015. This package requires member states to take specific measures to prioritize prevention, re-use, and recycling, according to the waste hierarchy. Recycling targets for municipal waste are as follows: (i) 55% by 2025, (ii) 60% by 2030, and (iii) 65% by 2035 [66].

The Italian Institute for Environmental Protection and Research (ISPRA) acts under the vigilance and policy guidance of the Italian Ministry for the Environment and the Protection of Land and Sea. Based on the data provided by this institute [67], Italian waste generation amounted to 30,116.6 ktons in 2016 (+2% from 2015), with a separated collection rate of 52.5% (+5% from 2015) (Table 2). Within the separate collected waste of the MSW, the OFMSW assumed a predominant position, representing about 41.2% of this mix, followed by paper and cardboard (20.3%), glass (11.7%), and plastic (7.8%).

Table 2. Statistical data in Italy regarding municipal solid waste (MSW) [67]. OFMSW—organic fraction
of municipal solid waste; SC—separated collection.

	2015 (1000*tons)	2016 (1000*tons)	2015 (kg/capita)	2016 (kg/capita)
Total MSW	29,524.3	30,116.6	486.7	497.1
Total SC of MSW	14,020.9	15,821.9	231.1	261.1
Total OFMSW	6071.5	6516.9	100.1	107.6

Moving from a national to more local perspectives, the present analysis concentrates on municipalities with populations greater than 200,000 (criteria fixed by ISPRA) (Figure 2). Table 3 presents the data for 15 municipalities that, in 2016, had a collective population of 9.9 million (16.4% of the Italian population) and produced 18.1% of the total national waste.



Figure 2. Map of 15 Italian municipalities.

Table 3. Statistical data in 15 Italian municipalities in 2016 [67].

	Total MSW (tons)	Collection Rate (%)	Total OFMSW (tons)	OFMSW (kg/capita)
Torino	427,741	42.1	53,312	60.1
Milano	679,278	57.6	139,740	103.4
Verona	138,787	55.3	25,223	98.0
Venezia	166,601	57.0	35,904	137.1
Padova	124,588	55.2	26,460	126.1
Trieste	94,104	40.2	8570	42.0
Genova	287,287	33.5	17,802	30.5
Bologna	223,011	46.0	22,834	58.8
Firenze	240,573	50.3	46,518	121.7
Roma	1,689,206	42.0	255,161	88.8
Napoli	519,421	31.3	58,017	59.8
Bari	191,328	36.7	16,309	50.3
Palermo	347,008	7.2	8458	12.6
Messina	113,442	11.2	2139	9.0
Catania	218,213	10.3	4457	14.2

3.2. Definition of Plant Size

A strategy is a plan of action taken by decision makers to achieve a final goal. Section 2 underlined that plant size assumes a key role in the economic analysis. Decision makers aim at reducing the environmental damage produced by inadequate waste management, and, at the same time, maximizing economic opportunities. Underestimating plant size prevents a share of the collected waste from being recovered, and its transfer causes higher emissions in the transport phase. In addition, this choice is not coherent with the principles of self-sufficiency and proximity (Directive 2008/98/CE). Overestimating the plant size can lead to economic losses when plant saturation is not achieved; consequently, the benefits of economies of scale must be balanced [68].

Biomethane plant size ($S_{biomethane}$) is defined according to three variables (see Equation (1)), namely: (i) the quantity of feedstock (Q_{OFMSW}), (ii) the unitary potential of biomethane (p_b^u), and (iii) the number of operative hours (n_{oh}).

$$S_{biomethane} = (Q_{OFMSW} * p_b^u) / n_{oh}$$
(1)

The amount of feedstock can be easily defined by the value reported in Table 3, but, at the same time, this table also reports different values for the separated collection rate (%SC). An increase in this last variable towards the final goal of 65% proposed by the European Commission also determines an increase in the collected OFMSW. The definition of the specific yields of SC for each fraction of waste was proposed by the authors of [69].

Four alternative scenarios are analyzed in this work, namely:

- 1. Scenario "As Is", in which Q_{OFMSW} is determined by the 2016 values.
- 2. Scenario "To Be—65%SC", in which the Q_{OFMSW} is calculated according to the recycling target of 65% and the %SC linked to the OFMSW is equal to the current mix.
- 3. Scenario "To Be—65%SC 40%OFMSW", in which the Q_{OFMSW} is defined according to the recycling target of 65% and the %SC of the OFMSW is equal to 40% (because this value is close to 41.2%, which was the 2016 average value %SC of the OFMSW in Italy). The percentage mix of separated waste is significantly diversified across the 15 municipalities, varying from 39.2% in the north to 45.7% in the south, and 41.7% in the central regions [67].
- 4. Scenario "To Be—110 kg/inhabitant", in which Q_{OFMSW} is evaluated as a function of the number of citizens. Some authors have assumed a yearly OFMSW production of 150 kg per inhabitant [70], considering good waste management practices. However, as discussed in Section 1, new policies that reduce the amount of waste generated also lower this value. A hypothetical reduction of 75% is considered, and, consequently, a yearly OFMSW production of approximately 110 kg per inhabitant is assumed.

For example, considering Torino as a case study, the Q_{OFMSW} is equal to 53,312 tons in the "As Is" scenario (Table 3). Considering that 427,741 tons of MSW equal to a SC of 65%, the amount of separated MSW is 278,032 tons.

Starting with this input, the Q_{OFMSW} is equal to 82,480 tons (i.e., almost 30% of 278,032 tons) and 111,213 tons (i.e., 40% of 278,032 tons) in the "To Be—65%SC" and "To Be—65%SC 40%OFMSW" scenarios, respectively. In fact, in the first situation, the %SC is approximately 30% (the 2016 value), while in the second situation it is 40%.

In these three scenarios, the amount of OFMSW production varies from 60 to 125 kg/inhabitant. Considering the current population of 886,837, the Q_{OFMSW} is equal to 97,552 tons (obtained by multiplying 886,837 inhabitants with 110 kg/inhabitant) in the "To Be—110 kg/inhabitant" scenario.

The biogas yield is mainly influenced by the specific nature of the OFMSW. The values of several feedstocks were proposed by the authors of [71]. Calculations of the biomethane yield typically assume a biogas composition of 60% methane, varying from 48 to 75 m³ biomethane/ton [72]; however, it was assumed to be equal to 78 m³ biomethane/ton by the authors of [73]. Considering the first industrial plant constructed in Italy, located in Montello (Bergamo), approximately 32 billion m³ biomethane are obtained by 600,000 tons of the OFMSW annually (approximately 55 m³ biomethane/ton) [74].

In addition, 100,000 tons/year of pig slurry is needed as feedstock for a 250 m³/h production capacity. The same result can be reached by 25,000 tons/year of the OFMSW or 18,000 tons/year of maize [67]. However, biogas and biomethane yields cannot be defined only as a function of the substrate. In fact, gas yields depend highly on the dry matter content, storage feedstock, and feedstock handling, and an exact estimate requires feedstock testing [75]. In this work, p_b^u is assumed equal to 75 m³ biomethane/ton [55].

Regarding the third variable, the literature typically assigns a value of 8000 operative hours, or approximately 333 days [45].

3.3. Policy Framework

With the biomethane sector not fully developed, a new decree for biomethane production was published by Italy's Economic Development Ministry (MISE) in March 2018. The decree provides for 4.7 billion € of incentives dedicated to plants, starting operations between 2018 and 2022 [44].

As advanced biofuels and biomethane have higher production costs than traditional fuels, the decree seeks to support biomethane producers with a premium to fill the cost gap. The incentives apply up to 1.1 billion m³ biomethane per year.

Transport fuel retailers have the obligation to sell a minimum amount of biofuel. This obligation can be satisfied in two ways, namely: (i) through selling the prescribed amount of biofuel in return for a corresponding amount of CICs, or (ii) through purchasing the obligation share of biofuels not injected into the market from producers or other obliged entities with a CIC surplus. A CIC is issued for each 10 GCal (single counting) of produced biomethane; considering that 1m³ CH₄ is equal to 8121 kcal, a CIC corresponds to about 1231 m³ CH₄. This calculation assumes that 1m³ CH₄ equals 0.68 kg under normal conditions (standard temperature of 273.15 K and pressure of 101.325 kPa).

The revenues for advanced biomethane are generated from two main sources, namely:

- 1. The payment of 375 € for a 10-year CIC; after this time, producers are only entitled to receive CICs at a price defined by the market. In addition, advanced biomethane, obtained by some substrates such as OFMSW, entitles producers to receive one CIC per 5 GCal (double counting).
- 2. Selling produced biomethane at the average price, weighed against the quantities registered in the virtual trading point (PTV) during the month of the sale, reduced by 5%.

Finally, an additional premium is offered when the producer is also the distributor of the methane. In fact, in this scenario, the value of the guaranteed CICs is increased by 20%. Additional data regarding the permits/authorization were proposed by the authors of [65], and these aspects are consistent and equal for all plants, regardless of their location and size.

3.4. Economic Model

The DCF analysis is a method of evaluating a project using the concept of the time value of money. It uses an incremental approach that considers cash inflows and outflows, and applies a cost opportunity of capital to aggregate several cash flows. NPV, which is the financial index used in this work, is the sum of the present values of the individual cash flows [42].

The subsidies, sale of biomethane, sale of food-grade CO₂ (through outsourcing), and the positive net income derived from treating the OFMSWs represent the revenue sources of a biomethane plant. Three phases typically characterize the biogas–biomethane chain, namely: (i) biogas production, (ii) upgrading, and (iii) compression and distribution. Investment, substrate, labor, transport, maintenance and overheads, depreciation for mechanical and electrical elements, electricity consumption, insurance, and the purchase of zeolite are the costs of a biomethane plant.

The following assumptions were used in this study:

- 1. The end specifications of gas (such as composition and pressure) were adjusted to its final use.
- 2. The CIC value was fixed for all 20 years. The lack of data did not permit us to conduct a statistical analysis of the CIC value. However, alternative values defined by the literature were used in addition to the baseline scenario.
- 3. The selling price of biomethane was fixed for all 20 years. In reality, this price could vary each month; but again, the scarcity of the available data did not enable us to conduct a statistical analysis. Alternative values were considered in addition to the baseline scenario.
- 4. The definition of an optimal size of biogas plant (S_{biogas}) was chosen in order to maximize the grade of saturation in the upgrading phase ($Q_{biomethane} \approx Q_{biomethane}^{nom}$).

A sensitivity analysis was aimed at verifying the results and reducing the subjectivity of the formulated hypothesis [76]. The model is as follows:

NPV =
$$\sum_{t=0}^{n} (I_t - O_t) / (1 + r)^t$$
 (2)

NPV/Size =
$$\left(\sum_{t=0}^{n} (I_t - O_t)/(1 + r)^t\right)/S_{\text{biomethane}}$$
 (3)

$$\sum_{t=0}^{DPB1} (I_t - O_t) / (1+r)^t = 0$$
(4)

$$I_{t} = R_{t}^{\text{subsidies}} + R_{t}^{\text{selling}} + R_{t}^{\text{CO}_{2}} + R_{t}^{\text{compost}} + R_{t}^{\text{OFMSW}}$$
(5)

$$R_{t}^{\text{subsidies}} = Q_{\text{biomethane}} * i_{\text{cic}}^{u} * c_{c} \forall t = 0...10$$

$$(6)$$

$$R_{t}^{\text{subsidies}} = Q_{\text{biomethane}} * p_{\text{cic}}^{u} \forall t = 11...n$$
(7)

$$i_{cic}^{u} = CIC / c_{cic}^{CH_4}$$
(8)

$$p_{cic}^{u} = CIC/c_{cic}^{CH_4}$$
(9)

$$R_t^{\text{selling}} = Q_{\text{biomethane}} * p_{\text{biomethane}} \quad \forall t = 0 \dots n$$
 (10)

$$p_{\text{biomethane}} = p_{\text{PTV}} * cf_{\text{gas}} * (1 - \% r p_{\text{PTV}})$$
(11)

$$R_t^{CO_2} = Q_{CO_2} * p_{CO_2}^u \quad \forall t = 0 \dots n$$

$$(12)$$

$$R_{t}^{compost} = Q_{compost,t}$$
(13)

$$R_{t}^{OFMSW} = Q_{OFMSW} * \left(R_{gross,t}^{OFMSW} - C_{t}^{OFMSW} \right) \forall t = 0 \dots n$$

$$(14)$$

$$0_{t} = C_{lcs,t}^{1^{\circ}s} + C_{lis,t}^{1^{\circ}s} + C_{lcs,t}^{2^{\circ}s} + C_{lis,t}^{2^{\circ}s} + C_{lcs,t}^{3^{\circ}s} + C_{lcs,t}^{3^{\circ}s} + C_{lcs,t}^{dig} + C_{lis,t}^{dig} + C_{l,t}^{dig} + C_{s,t} + C_{ts,t} + C_{m_{ko,t}}^{1^{\circ}s} + C_{df,t}^{1^{\circ}s} + C_{e,t}^{1^{\circ}s} + C_{i,t}^{1^{\circ}s} + C_{m_{ko,t}}^{2^{\circ}s} + C_{df,t}^{2^{\circ}s} + C_{e,t}^{2^{\circ}s} + C_{o,t}^{2^{\circ}s} + C_{ts,t}^{2^{\circ}s} + C_{ts,t}^{2^{\circ}s} + C_{o,t}^{2^{\circ}s} + C_{ts,t}^{2^{\circ}s} + C_{ts,t}^{2^{\circ}s} + C_{ts,t}^{2^{\circ}s} + C_{o,t}^{2^{\circ}s} + C_{o,t}^{2^{\circ}s} + C_{ts,t}^{2^{\circ}s} + C_{ts,t}^{2$$

$$C_{inv}^{1^{\circ}s} = C_{inv}^{u,1^{\circ}s} * S_{biogas}$$
(16)

$$C_{lcs,t}^{1^{\circ}s} = C_{inv}^{1^{\circ}s} / n_{debt} \quad \forall t = n_{(ba)} . n_{debt} - 1$$
(17)

$$C_{lis,t}^{1^{\circ}s} = \left(C_{inv}^{1^{\circ}s} - C_{lcs,t}^{1^{\circ}s} \right) * r_{d} \quad \forall t = n_{(ba)} \cdot n_{debt} - 1$$
(18)

$$C_{\rm inv}^{2^{\circ}s} = C_{\rm inv}^{u,2^{\circ}s} * S_{\rm biomethane}$$
(19)

$$C_{lcs,t}^{2^{\circ}s} = C_{inv}^{2^{\circ}s} / n_{debt} \quad \forall t = 0 \dots n_{debt} - 1$$
(20)

$$C_{lis,t}^{2^{\circ}s} = \left(C_{inv}^{2^{\circ}s} - C_{lcs,t}^{2^{\circ}s}\right) * r_{d} \ \forall t = 0 \dots n_{debt} - 1$$
(21)

$$C_{\rm inv}^{3^{\circ}{\rm s}} = C_{\rm inv}^{\rm com} + C_{\rm inv}^{\rm dis}$$
⁽²²⁾

$$C_{\text{lcs},t}^{3^{\circ}\text{s}} = C_{\text{inv}}^{3^{\circ}\text{s}} / n_{\text{debt}} \quad \forall t = 0 \dots n_{\text{debt}} - 1$$
(23)

$$C_{lis,t}^{3^{\circ}s} = \left(C_{inv}^{3^{\circ}s} - C_{lcs,t}^{3^{\circ}s}\right) * r_{d} \forall t = 0 \dots n_{debt} - 1$$
(24)

$$C_{lcs,t}^{dig} = C_{inv}^{dig} / n_{debt} \ \forall t = 0 \dots n_{debt} - 1$$
(25)

$$C_{\text{lis},t}^{\text{dig}} = \left(C_{\text{inv}}^{\text{dig}} - C_{\text{lcs},t}^{\text{dig}}\right) * r_d \quad \forall t = 0 \dots n_{\text{debt}} - 1$$
(26)

$$C_{l,t} = C_l^{u,a} * n_{op} \forall t = 0 \dots n$$

$$(27)$$

$$C_{s,t} = C_s^u * Q_{OFMSW} \ \forall t = 0 \dots n$$
(28)

$$C_{ts,t} = C_{ts}^{u} * Q_{OFMSW} \ \forall t = 0 \dots n$$
⁽²⁹⁾

$$C_{mo,t}^{1^{\circ}s} = p_{mo}^{1^{\circ}s} * C_{inv}^{1^{\circ}s} \ \forall t = 0 \dots n$$
(30)

$$C_{df,t}^{1^{\circ}s} = p_{df} * C_{lcs,t}^{1^{\circ}s} \forall t = 0...n$$
(31)

$$C_{e,t}^{1^{\circ}s} = c_e^{u,1^{\circ}s} * Q_{\text{biogas}} * p_e \ \forall t = 0 \dots n$$

$$(32)$$

$$C_{i,t}^{1^{\circ}s} = p_{i} * C_{inv}^{1^{\circ}s} \forall t = 0...n$$
(33)

$$C_{mo,t}^{2^{\circ}s} = p_{mo}^{2^{\circ}s} * C_{inv}^{2^{\circ}s} \ \forall t = 0 \dots n$$
(34)

$$C_{df,t}^{2^{\circ}s} = p_{df} * C_{lcs,t}^{2^{\circ}s} \ \forall t = 0 \dots n$$
(35)

$$C_{e,t}^{2^{\circ}s} = c_e^{u,2^{\circ}s} * Q_{\text{biogas}} * p_e \ \forall t = 0 \dots n$$

$$(36)$$

$$C_{i,t}^{2^{\circ}s} = p_i * C_{inv}^{2^{\circ}s} \ \forall t = 0 \dots n$$
(37)

$$C_{z,t}^{2^{\circ}s} = p_z^{u} * Q_z \quad \forall t = 0 \dots n$$
(38)

$$C_{o,1}^{com} = C_o^{com} \tag{39}$$

$$C_{o,1}^{dis} = C_o^{dis} \tag{40}$$

$$C_{gv,t+1} = C_{gv,t} * (1 + inf) \forall t = 0...n$$

$$\tag{41}$$

$$C_{tax,t} = CF_t * p_{tax} \tag{42}$$

$$Q_{biogas}^{nom} = S_{biogas} * n_{oh} * \% CH_4$$
(43)

$$Q_{biogas} = Q_{biogas}^{nom} * (1 - l_{bs})$$
(44)

$$Q_{biomethane}^{nom} = S_{biomethane} * n_{oh}$$
(45)

$$Q_{biomethane} = Q_{biogas} * (\% CH_4) * (1 - l_{us}) * r_{bm}$$
(46)

$$Q_{CO_2} = S_{biogas} * n_{oh} * (\% CO_2) * cf_{CO_2} * r_{CO_2}$$
(47)

3.5. Input Assumptions

The robustness of an economic analysis is determined by the choice of input variables. For this reason, this work opted to use the values reported in the literature (Table 4), and, regarding subsidies, followed the inputs reported in the decree. The opportunity cost of the capital measured the return of a project with the same risk level, and was assumed to be equal to 5%. The lifetime of the project was hypothesized to equal the period of the subsidies provided in the previous version of the decree, which was 20 years. Year 0 was fixed in 2018, and 6 months was defined as the construction time of the plant. A pressure swing adsorption (PSA) technique was considered in the upgrading phase. Regarding the unitary value of the subsidy (i_{cic}^{u}), the following value was used:

$$i_{cic}^{u} = 375/1231 = 0.305 \ /m^{3}$$
 single counting (48)

$$i^{u}_{cic} * c_c = 0.305 * 2 = 0.610 / m^3$$
 double counting (49)

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Variable	Value	Reference
c	2	[44]
CH_4	1231 m ³	[65]
cic u,1°s e	0.13 kWh/m ³	[77]
1.7.5	0.29 kWh/m ³	[72]
e rcom	53,000€	[78]
ʻinv ,dig	680,000–2,250,000	
inv	€	[79]
dig	237,500€	[80]
iny u,1°s	4000–5000 €/kW	
inv		[27]
su,2°s	2250–6300	[27]
inv u,a	€/(m ³ /h)	[(0]
l com	25,000 €/y	[63]
0	47,000 €/y	[78]
dig o	45,000–440,000	[79]
dis	€/y 20.000 €/m	
0 U	20,000 €/y	[80]
3 OFMSW	0€ 10 C/L	[81]
dis o u S OFMSW t t ts	49 €/ton	[27]
ts	2–3 €/ton	[81]
CO ₂	1.84 kg/m^3	[82]
gas	0.0105 MWh/m ³	[63]
IC	375€	[44]
c	0.305 €/m ³	[44]
ıf	2%	[27]
5	6%	[80]
s	1.5%	[80]
	20 y	[27]
debt	15 y	[83]
oh	8000 h	[83]
pp	6	[81]
6	10 y	[44]
1)	see Section 3.2	[25]
biomethane	see Section 3.5	[44]
1 zic	see Section 3.4	[27]
20 ₂	10 €/ton	[79]
compost	50 €/ton	[84]
lf	10%	[82]
2	0.13€/kWh	[79]
	1%	[84]
l°s mo	20%	[85]
2°š mo	10%	[84]
PTV	see Section 3.5	[72]
ax	40%	[63]
1	800 €/ton	[81]
compost	2000–15,000 tons	[82]
OFMSW	see Section 3.2	
z	50 ton/y	[81]
	5%	[27]
m	97%	[42]
2O ₂	97%	[81]
l	3%	[83]
OFMSW gross,t	70 €/ton	[27]
gross,t piogas	150–1500 kW	[]
piomethane	$50-500 \text{ m}^3/\text{h}$	
CH ₄	60%	[86]
CO_2	47%	[87]
rp _{PTV}	5%	[44]

Table 4. Input assumptions.

Concerning the selling price of biomethane ($p_{biomethane}$), some data were available (p_{PTV}), namely: from 19.44 €/MWh (February 2019) to 27.78 €/MWh (July 2018) [43]. The conversion factor (cf_{gas}) was 1 m³ = 0.0105 MWh. The value used in this work was the average of the following:

$$p_{\text{biomethane}} = 0.25 \ / \text{m}^3 \text{ average value}$$
 (50)

The unitary investment cost of both the AD and the upgrading phases decreased with the increasing plant size. The biogas production varied from $5300 \notin kW$ (for a 150 kW plant) to $4000 \notin kW$ (for a 1500 kW plant). Biomethane production, in contrast, ranged from $6300 \notin (m^3/h)$ for a 50 m³/h plant, to $2250 \notin (m^3/h)$ for a 500 m³/h plant.

3.6. Environmental Analysis

Regarding the environment, the GHG emissions of the vehicle power train systems were recorded, from the "well to wheel." Section 2 defined the reduction of GHG emissions as 119 gCO₂.eq/km when BIO-CNG was used as an alternative to CNG. In fact, the actual GHG emissions when BIO-CNG was used were 5 gCO₂.eq/km (GHG_{BIOCNG}), relative to those from fossil fuels, which were 124 gCO₂eq/km (GHG_{CNG}) [50]. The environmental analysis is focused only on the reduction of GHG emissions obtained by the CNG vehicles fueled by biomethane, according to the authors of [19,50,51,55,63].

It was hypothesized that an NGV travelled 15,000 km per year (n_{km}) [63], and the reduction of GHG emissions of a NGV (uRGHG_{NGV}) could be estimated as 1785 kgCO₂.eq/y. The number of NGVs fueled by BIO-CNG (n_{NGV}) was calculated as the ratio between the amount of produced biomethane $(Q_{biomethane})$ and the annual NGV consumption (c_{NGV}) . For example, a plant with a unitary capacity $(1 \text{ m}^3/\text{h})$ produced about 7760 m³ of biomethane, and, considering an annual methane consumption of 1100 m³ [27], about seven NGVs would be powered by BIO-CNG.

The environmental benefit of a biomethane plant was obtained by multiplying the unitary value of the emissions reduction with the number of NGVs. For example, the reduction of GHG emissions obtained by the NGVs fueled by biomethane (RGHG_{NGV}) was found to be12.6 tCO₂-eq/y in a plant of $1 \text{ m}^3/\text{h}$.

$$RGHG_{NGV} = (GHG_{CNG} - GHG_{BIOCNG}) * n_{km}$$
(51)

$$uRGHG_{NGV} = 119 * 15,000 = 1785 kgCO_2 eq/y$$
 (52)

$$n_{\rm NGV} = (S_{\rm biomethane} * n_{\rm oh} * r_{\rm bm}) / c_{\rm NGV}$$
(53)

$$n_{\rm NGV} = (1*8000*0.97)/1100 = 7 \,\rm NGVs \tag{54}$$

$$RGHG_{NGV} = n_{NGV} * uRGHG_{NGV}$$
(55)

$$RGHG_{NGV} = 7.05 * 1785 = 12.6tCO_2 eq/y$$
(56)

4. Results

This section begins by identifying the criteria used to define the minimum plant size that can fulfil the profitability condition (i.e., NPV >0). Ten cases relating to plants of different sizes, ranging from 50 m³/h to 500 m³/h, are considered (Section 4.1). Following this, a sensitivity analysis is conducted to strengthen the obtained results (Section 4.2), and the scenario in which the biomethane producer is also the methane distributor of the methane is examined (Section 4.3). The potential productivity of biomethane and the associated profits are reported in Sections 4.4 and 4.5, respectively. Also, alternative scenarios are proposed in Section 4.6.

4.1. Profitability Analysis of Biomethane Plants—Baseline Scenario

The profitability of the biomethane plants powered by OFMSW was confirmed in seven of out of the ten cases proposed in Figure 3. The NPV was positive for the 200 m^3/h plant and greater than that

of the 150 m³/h plant proposed in the literature. However, this result depended on the specific input data used (see Section 2).

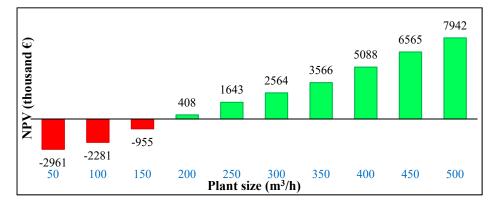


Figure 3. Profitability analysis of biomethane plants (thousand €)—baseline scenario. NPV—net present value.

Although the NPV measured the plant profitability, the index was not able to consider different size values. For this reason, considering the different plant capacities examined in this work, the ratio of NPV/size was useful for comparison. NPV/size was coherent with NPV, and was obtained by dividing this indicator with the plant size; thus, all of the case studies were homogeneous, because they were expressed in terms of \in for m³/h.

The profits obtained were extremely consistent, varying from $2040 \notin$ for m³/h (200 m^3 /h plant), to $15,884 \notin$ for m³/h (500 m^3 /h plant). This was determined by two main variables, namely: (i) economies of scale and (ii) the structure of subsidies, which did not differentiate between plant size.

The use of OFMSW as a substrate determined the profitability of the biomethane plants because of the effect of double counting, in which the value of the corrective coefficient (c_c) is equal to 2. Subsidies played a key role in the discounted cash inflows (approximately 39%; Figure 4). As the MSW treatment is a paid service, its value exceeded the necessary costs for pre-treatment (equal to 21 \notin /ton; Table 4). The use of organic waste provided additional income to the biogas–biomethane plants, via a tariff paid for the treatment of waste; in fact, the tariff payment for the OFMSW treatment was paid to the plant owner by the local municipality [88,89]. The contribution of this net revenue was significant, representing approximately one quarter of the total (23%). Of course, willingness to pay for separate waste collection depends on both economic and non-economic motivations [90]. The sale of biomethane influenced about one fifth of the total (20%).

Regarding the discounted cash outflows, the percentage weight of the investment costs was significantly lower than that of the operative costs (15%). The investment costs consisted of both a loan capital share and a loan interest share, because it was hypothesized that these costs would be covered by third party funds. The costs were subdivided for a single phase of biomethane production as follows: (i) AD, (ii) upgrading, and (iii) compression and distribution. The following operative costs were analyzed: substrate, labor, transport, maintenance and overheads, depreciation for mechanical and electrical devices, electricity consumption, and insurance. Also, some items were specific to biogas production and upgrading. Among these, the maintenance and overhead costs (in particular during biogas production) were the most relevant (41%). The average values of the costs linked with several plant sizes are reported in Figure 4, with a null value for the substrate (Table 4). Starting with these items, the level cost for the production of 1 Nm³ of biomethane was obtained as the ratio between the sum of the discounted cash outflows and the biomethane production. Its value varied from $0.63 \ embed{m}^3$ (50 m³/h plant). These values were greater than the values registered for the production of natural gas; for this reason, it would be necessary to apply a subsidy. This renewable resource is better able to tackle climate change than fossil fuels [76,91].

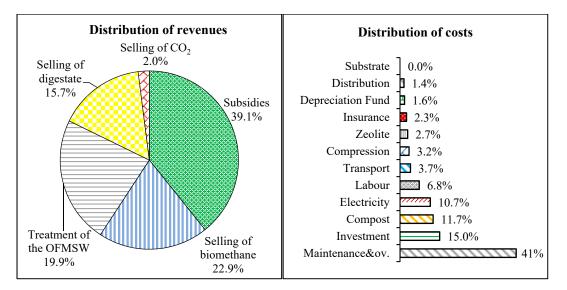


Figure 4. Distribution of discounted cash flows (%)—baseline scenario. OFMSW—organic fraction of municipal solid waste.

As payback time (PBT) is an inappropriate indicator for measuring the economic feasibility of a project, this work used discounted payback time (DPBT), which considers the real value of money. DPBT was less than 1 year in several plants (from 250 to 500 m³/h), and equal to 1 year in a 200 m³/h plant. These results were coherent with NPV, and were encouraged mainly for two reasons, namely: (i) the assumption of third party funding with investment costs spread over the debt period, and (ii) the lower weight of investment costs relative to operative costs. A main challenge of DPBT is that it does not take into account what happens after payback. Specifically, some cash flow values may be negative towards the end of the debt period, because the economic model assumes that the costs will increase during the period because of inflation, while the revenues will not be subjected to such variation.

4.2. Profitability Analysis of Biomethane Plants—Alternative Scenarios

The economic results obtained in the baseline scenario are linked to the input variable assumptions. All of the business plans demonstrate dynamic aspects, and, consequently, variability in the results is generally found. This issue is typically resolved through a sensitivity analysis conducted on the critical variables. The critical variables were defined according to the new version of the decree and the existing literature [63]. Optimistic (opt) and pessimistic (pes) scenarios were proposed for the three strategic variables, as follows:

- The CIC was not assumed to be fixed for the lifetime of the biomethane plant. After 10 years, the unitary value of the subsidy (p^u_{cic}) was considered equal to 0.244 €/m³ with a CIC of 300 € (scenario R^{sub}_{pes}), and 0.406 €/m³ with a CIC of 500 € (scenario R^{sub}_{opt}).
- The selling price of the biomethane (p_{biomethane}) was considered fixed for 20 years, but different in the two scenarios, namely: 0.20 €/m³ (scenario R^{sel}_{pes}) and 0.30 €/m³ (scenario R^{sel}_{opt}).
- The price paid for the MSW recovery was not the same for all of the municipalities, and the mix of OFMSW determined a different cost of treatment. The net revenues (R^{OFMSW}_{gross} C^{OFMSW}) were modified to 15 €/ton (scenario R^{ofm}_{pes}) and 30 €/ton (scenario R^{ofm}_{opt}).

Table 5 outlines the 60 alternative scenarios obtained by the combination of the ten plant sizes, three critical variables, and two estimates of value. The ratio between NPV and plant size is shown in Figure 5. NPV was chosen as a measurement because of its wide use for evaluating the profitability of biomethane plants [27,45,61,92,93].

	R ^{sub} _{pes}	R ^{sub} opt	R ^{sel} _{pes}	R ^{sel} opt	R ^{ofm} _{pes}	R ^{ofm}
50 m ³ /h	-3079	-2764	-3284	-2713	-3710	-2348
100 m ³ /h	-2518	-1888	-2777	-1789	-3098	-1147
150 m ³ /h	-1311	-366	-1533	-377	-2868	473
200 m ³ /h	-67	1115	-363	1153	-863	2172
250 m ³ /h	1082	2444	729	2511	107	3733
300 m ³ /h	1913	3483	1494	3586	777	5005
350 m ³ /h	2849	4581	2352	4733	1532	6331
400 m ³ /h	4312	6131	3735	6373	2824	8131
450 m ³ /h	5736	7657	5079	7935	4080	9908
500 m ³ /h	7073	9126	6337	9434	5249	11,626

Table 5. Profitability analysis of biomethane plants (thousand €)—alternative scenarios.

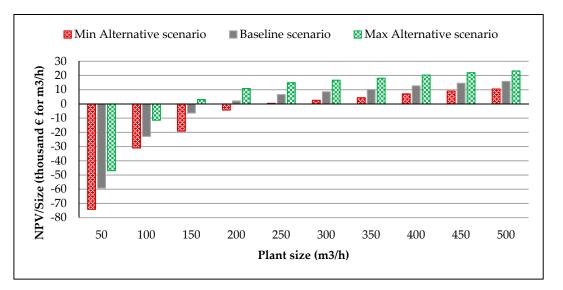


Figure 5. NPV/size (thousand € for m³/h)—baseline and alternative scenarios.

The alternative scenarios strengthened the obtained results. NPV was always positive for a 250 m³/h plant, and this became the new minimum size, even considering the negative scenarios. In fact, for the negative scenarios (R_{pes}^{sub} , R_{pes}^{sel} , and R_{pes}^{ofm}), the profitability of a 200 m³/h plant was not verified. However, economic feasibility was reached in a 150 m³/h plant in the R_{opt}^{ofm} scenario.

A probability value was not assigned to each scenario, and a comparison was not possible because the variation of critical variables was not uniform. Investors rely on the development of these scenarios, because varying assumptions with respect to the input data define economic performance. In particular, in our analysis, the minimum and maximum values were linked to the variation in net revenues from the OFMSW treatment. For example, in a 200 m³/h plant, the NPV varied from -4315 for m³/h (-4.3 K€ for m³/h) to 10,860 € for m³/h (10.9 K€ for m³/h). The structure of the subsidies, which remained consistent for all of the plant sizes, characterized these results. In fact, maximum profits were obtained with larger sizes of upgrading plants, but only when they operated at full saturation. For this reason, the definition of plant size must be linked to the number of substrates treated, and the opportunity to manage multiple substrates.

4.3. Joint Implementation between a Biomethane Producer and Methane Distributor

The new version of the decree recognizes a special premium for investors that both produce biomethane and distribute methane. In this scenario, the CIC value is increased by 20% for the first 10 years. Thus, the previous model was integrated as follows:

$$R_{t}^{\text{subsidies}} = Q_{\text{biomethane}} * i_{\text{cic}}^{u} * c_{\text{c}} * c_{\text{pos}} \forall t = 0...10$$
(57)

$$R_{t}^{\text{subsidies}} = Q_{\text{biomethane}} * p_{cic}^{u} \forall t = 11...ns$$
(58)

$$p_{\text{biomethane}} = p_{\text{pos}} \tag{59}$$

$$0_{t} = C_{lcs,t}^{1^{\circ}s} + C_{lis,t}^{1^{\circ}s} + C_{lcs,t}^{2^{\circ}s} + C_{lis,t}^{3^{\circ}s} + C_{lis,t}^{3^{\circ}s} + C_{lcs,t}^{4^{\circ}s} + C_{lis,t}^{4^{\circ}s} + C_{lis,t} + C_{lis,t} + C_{lis,t} + C_{lis,t} + C_{lis,t} + C_{lis,t} + C_{lis,t}^{2^{\circ}s} + C_{lis,t}^{$$

 $+C_{tax,t}$

$$C_{\rm inv}^{4^{\circ}s} = C_{\rm inv}^{\rm pos} \tag{61}$$

$$C_{lcs,t}^{4^{\circ}s} = C_{inv}^{4^{\circ}s} / n_{debt} \quad \forall t = n_{(ba)} \cdot n_{debt} - 1$$
(62)

$$C_{lis,t}^{4^{\circ}s} = \left(C_{inv}^{4^{\circ}s} - C_{lcs,t}^{4^{\circ}s}\right) * r_{d} \forall t = n_{(ba)} \cdot n_{debt} - 1$$
(63)

$$C_{o,1}^{\text{pos}} = C_o^{\text{pos}} \tag{64}$$

where pos = point of sale; $c_{pos} = corrective coefficient (point of sale)$; $p_{pos} = price of biomethane (point of sale); C_{inv}^{pos} = investment cost (point of sale); and <math>C_o^{pos} = operative cost (point of sale)$.

In the new decree, the unitary value of the subsidy increases by two corrective coefficients, as follows: the double counting (c_c) coefficient and the corrective coefficient (c_{pos} equal to 1.2), because of the application of a distribution premium [44].

$$i_{cic}^{u} * c_{c} * c_{pos} = 0.305 * 2 * 1.2 = 0.732 / m^{3}$$
 double counting (65)

Consequently, this value is equal to 0.732 €/m^3 for the first 10 years, after which it is assumed equal to 0.305 €/m^3 . The input data regarding the point of sale were chosen according to a report elaborated by the authors of [94]. The investment and operative costs of the distribution plant were assumed equal to $600,000 \text{ € } (C_{inv}^{pos} \text{ for both } 50 \text{ and } 100 \text{ m}^3/\text{h} \text{ plants equal to } 500,000 \text{ € }) and <math>1680 \text{ €/(m}^3/\text{h}) (C_o^{pos})$. The price to the final consumer of the natural gas (net value added tax and excise) replaced the selling price of biomethane, and was assumed equal to $0.529 \text{ €/m}^3 (\text{p}_{pos})$ in the R^{pos} scenario. However, in this case, a sensitivity analysis was also conducted. Two alternative scenarios were analyzed, in which p_{pos} was assumed to be equal to 0.40 €/m^3 (scenario R_{pes}^{pos}) and 0.60 €/m^3 (scenario R_{opt}^{pos}) (Table 6).

Table 6. Profitability analysis of biomethane plants (thousand €)—point of sale scenarios.

	R ^{pos}	R ^{pos} _{pes}	R ^{pos} _{opt}
50 m ³ /h	-3011	-3650	-2659
100 m ³ /h	-1862	-3068	-1296
150 m ³ /h	-481	-1972	340
200 m ³ /h	1147	-966	2019
250 m ³ /h	2649	242	3898
300 m ³ /h	3891	1060	5366
350 m ³ /h	5390	2124	7065
400 m ³ /h	7015	3335	8848
450 m ³ /h	8679	4626	10,657
500 m ³ /h	10473	6071	12,607

This business model confirmed the trend of profitability proposed in the previous analysis. In particular, NPV was always positive for a 250 m³/h plant. In addition, a lower selling value of biomethane determined a negative NPV in a 200 m³/h plant (scenario R_{pes}^{pos}). However, the profitability was verified for a 150 m³/h plant, considering a greater selling value of biomethane (scenario R_{opt}^{pos}). The application of a premium determined an increase in revenues, which always exceeded the costs. This had the effect of increasing the economic performance. For example, in a 250 m³/h plant, NPV

increased 1006 K \in in comparison to the baseline scenario, and 3648 K \in when the selling price of the biomethane was assumed to be equal to $0.60 \notin m^3$. However, an opposite situation was verified in the R_{pes}^{pos} scenario, with a decrease of NPV equal to 1401 K \in . The DPBT was less than 1 year, in line with the previous results.

4.4. The Calculation of Potential Biomethane Productivity

Following the economic descriptions, this section provides details on potential biomethane productivity. Table 7 outlines the potential amount of the OFMSW to be treated, according to the four scenarios defined in Section 3.2.

Scenarios	As Is	To Be—65%SC	To Be—65%SC 40% OFMSW	To Be—110 kg/Inhabitant
Torino	53,312	82,480	111,213	97,552
Milano	139,740	157,689	176,612	148,672
Verona	25,223	29,851	36,085	28,309
Venezia	35,904	41,480	43,316	28,810
Padova	26,460	31,518	32,393	23,081
Trieste	8570	14,075	24,467	22,466
Genova	17,802	35,196	74,695	64,196
Bologna	22,834	32,486	57,983	42,720
Firenze	46,518	62,339	62,549	42,049
Roma	255,161	400,301	439,194	316,084
Napoli	58,017	128,481	135,049	106,721
Bari	16,309	29,397	49,745	35,662
Palermo	8458	77,453	90,222	74,111
Messina	2139	12,388	29,495	26,066
Catania	4457	28,911	56,735	34,473
Total	720,904	1,164,045	1,419,753	1,090,970

Table 7. Potential amount of the OFMSW conferred to biomethane plants (tons).

As expected, the amount of OFMSW changed significantly across the four scenarios. Certainly, the values proposed in the "To Be" scenarios were greater than those of the "As Is" scenario (equal to 721 ktons). This means that a substantial amount of waste became unsorted waste and some entered landfills, violating the good practices of the waste hierarchy. In this way, these municipalities did not maximize the recycled/recovered quantity. Only Milano, Verona, Venezia, and Padova had a recycling rate greater than 55%; at the same time, this value was lower than 12% in Palermo, Messina, and Catania (see Table 2). For this reason, two scenarios, "To Be-65%SC" and "To Be-65%SC 40%OFMSW," were constructed. The target recycling rate of 65% was fixed, and an increase in the sorted waste determined an increase of all of the components in this mix (e.g., the OFMSW, paper, glass, and plastic). The total amount of OFMSW in "To Be-65%SC" was 1164 ktons, which was lower than 1420 ktons ("To Be—65%SC 40%OFMSW"). In fact, the weight of the OFMSW in the collected waste mix in the "As Is" scenario was lower than 40% in all of the municipalities, varying from 17% in Messina to 39% in Padova. It was not possible to compare these two scenarios from an environmental perspective. However, the same cannot be said for the "To Be—110 kg/inhabitant" scenario (equal to 1091 ktons). In fact, another good practice of waste management is reducing generated waste. A defined value (in this case 110 kg/inhabitant) could counter this approach.

According to Equation (1), it was possible to calculate the potential biomethane productivity expressed in m³ per hour (Table 8). In this way, the value could be interpreted as the biomethane plant size.

Scenarios	As Is	To Be—65%SC	To Be—65%SC 40% OFMSW	To Be—110 kg/Inhabitant
Torino	500	773	1043	915
Milano	1310	1478	1656	1394
Verona	236	280	338	265
Venezia	337	389	406	270
Padova	248	295	304	216
Trieste	80	132	229	211
Genova	167	330	700	602
Bologna	214	305	544	401
Firenze	436	584	586	394
Roma	2392	3753	4117	2963
Napoli	544	1205	1266	1001
Bari	153	276	466	334
Palermo	79	726	846	695
Messina	20	116	277	244
Catania	42	271	532	323
Total	500	773	1043	915

Table 8. Potential biomethane productivity (m³/h).

4.5. Calculation of Potential Biomethane Profits—Baseline Scenario

The decisional criteria used in this work were as follows:

- A plant was built only if economic feasibility was reached, according to plant size. As demonstrated in Figure 3, a 200 kt plant was the minimum size.
- The pant size was defined according to Table 8, with allowance for a slight approximation by excess.
- The plant size varied from 200 kt to 500 kt, in multiples of 50 kt.
- The NPV obtained in the baseline scenario was considered the economic profit of reference.

Table 9 presents the number of biomethane plants, the definition of the plant size, and the calculation of the economic profit for each municipality. For example, there were four 500 m³/h plants and one 400 m³/h plant in Rome in the "As Is" scenario, and eight 500 m³/h plants constructed in two "To Be" scenarios. The main results are reported in Table 10. Fifteen municipalities could achieve profit in an "As Is" scenario characterized by eight 500 m³/h plants; two 250 m³/h plants; and one 200, 300, 350, 400, and 450 m³/h plant. This determined a biomethane productivity equal to 6200 m³/h; considering an operational period of 8000 h, the production of biomethane was equal to 49.6 million m³ (approximately 5% of the amount supported by incentives (1100 million m³)—see Section 3.3). The advantages of biomethane could be explained from an environmental perspective (about 78.1 thousand tons of CO₂.eq were avoided annually, and, consequently, 1.6 million tons of CO₂.eq were avoided over a period of 20 years) and an economic perspective (as the value of profits were 85 million \notin during the project lifetime).

The results of this work underline that biomethane plants support the attainment of several goals, as follows:

- The OFMSW produced by the citizens can pose a challenge to the ecosystem. In fact, the incorrect
 management of these flows towards landfills or long-distance transport to a limited number of
 plants increases GHG emissions. Thus, the biogas-biomethane chain reduces environmental
 damage [95,96].
- The biomethane obtained by OFMSW transforms this waste into a resource. Several economic opportunities are generated and the profitability of plants undertaking this processing can reduce the tax bills paid by citizens. In fact, economic feasibility is also obtained when net revenues are reduced in the treatment of the OFMSW [27,61]. The final goal is to stimulate the citizens

to separate waste in order to lower their waste collection bill. Should this occur, the revenues received by the municipalities would be reduced, and, consequently, the payments made by the municipalities to the OFMSW treatment plants would also be reduced.

- Biomethane produced by OFMSW increases the share of renewable energy in the transport sector. This goal is urgent. In fact, in a recent report issued by the European Commission, Italy was noted to have already achieved the EU target for renewables in the electricity sector, while being far from achieving their target in the transport sector. In addition, there is a serious delay regarding the national sub-target for advanced biofuels [94].
- Italy has a great number of NGVs (1,001,614 in 2016, about 77% of the European market) [97] and a developed biogas market. At the same time, the biomethane market is not well developed. NGVs can be fueled by this renewable resource and, in this way, the levels of imported gas can be lessened, reducing geopolitical risk [63,83]. In the "To Be—65%SC" scenario, the quantity of 85.2 million m³ of biomethane was able to fuel 77,455 NGVs. Considering that each NGV can travel 10,000 or 20,000 km, the number of NGVs changed to 116,712 and 58,356, respectively. Consequently, the number of NGVs fueled by biomethane changed as a function of both of the "To Be" scenarios and the number of km traveled. The values examined in this work ranged from 55,068 to 140,822 NGVs.

The analysis of this work was limited to 15 municipalities in Italy. There were 26 and 28 plants that were hypothesized in the "To Be—65%SC" and "To Be—65%SC 40%OFMSW" scenarios, respectively. The annual production of biomethane was found to vary from 85.2 to 102.8 million m³, equal to about 8% to 10% of the funds made available in the new version of the decree. The economic benefits were found to be extremely consistent, varying from 145 to 187 million \in , with an annual reduction of GHG emissions ranging from 134 to 162 million kgCO₂-eq.

Scenarios	Scenarios As Is		To Be—65%SC			To Be—65%SC 40% OFMSW		To Be—110 kg/inhabitant	
	Size	NPV	Size	NPV	Size	NPV	Size	NPV	
Torino	500	7942	500	9585	500 [2]	15,884	500	13,030	
			+250				+400		
Milano	500 [2] +300	18,448	500 [3]	23,826	500 [3]	23,826	500 [2] +400	20,972	
Verona	250	1643	300	2564	350	3566	250	1643	
Venezia	350	3566	400	5088	400	5088	250	1643	
Padova	250	1643	300	2564	300	3564	200	408	
Trieste	0	0	0	0	250	1643	200	408	
Genova	0	0	350	3566	500 +200	8350	500	7942	
Bologna	200	408	300	2564	500	7942	400	5088	
Firenze	450	6565	500	7942	500	7942	400	5088	
Roma	500 [4] +400	36,856	500 [7] +250	57,237	500 [8]	63,536	500 [5] +450	46,275	
Napoli	500	7942	500 [2]	16,292	500 [2] +250	17,527	500 [2]	15,884	
			+200		+250				
Bari	0	0	300	2564	450	6565	350	3566	
Palermo	0	0	500	9585	500	11,508	500	8350	
			+250		+350		+200		
Messina	0	0	0	0	300	2564	250	1643	
Catania	0	0	250	1643	500	7942	300	2564	
		Size (m ³	/h); NPV (th	ousand €); [1	n] = number	of plants			

Table 9. Definition of plant size and calculation of economic profit (thousand €)—baseline scenario. NPV—net present value.

As Is	To Be—65%SC	To Be—65%SC 40% OFMSW	To Be—110 kg/inhabitant
		Number of biomethane plants	
15	26	29	25
		Biomethane productivity (m ³ /h)	
6200	10,650	12,850	10,050
		Biomethane (million m ³ /y)	
49.6	85.2	102.8	80.4
		Economic benefits (million €)	
85	145	187	135
	Envir	onmental benefits (thousand tCO ₂₋ e	q/y)
78	134	162	127

	Table 10. Mai	n results in the	15 municip	alities from	managing the	OFMSW-	-baseline scenario.
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4.6. Calculation of Potential Biomethane Profits—Alternative Scenarios

The decisional process was influenced by the definition of the minimum plant size. The economic results of the alternative scenarios reported in Table 5 underlined that this value varied from 150 to 200 m³/h plant. At the same time, the economic performance of the plants ranged significantly, as determined by the total economic benefits linked to the 15 municipalities. This work was based on two estimates. The first considered that citizens might be opposed to the realization of large plants; for this reason, the maximum size considered was 500 m³/h. The second concerned the assumption that the value of the potential feedstock would be calculated for each municipality. However, flows of waste coming from nearby cities could be managed for economic reasons, with respect to the environmental principle of proximity. Moreover, other substrates (as by-products for which double counting would apply) could be employed as a mix with OFMSW in these plants.

The legislation defines a premium for investors that both produce biomethane and distribute methane. Table 6 defined the minimum size as 200 m³/h, and the NPV increased at all of the sizes. Table 11 presents the calculations of the economic profit, and Figure 6 illustrates the economic benefits of the alternative scenarios. The premium provided for a methane distributor did not modify the biomethane productivity in several scenarios. Instead, the total economic benefits varied significantly.

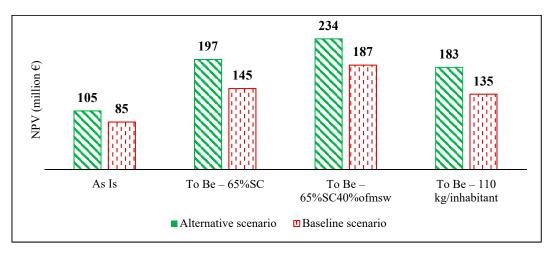


Figure 6. Economic benefits (million €) in the 15 municipalities from managing the OFMSW—alternative scenario (joint implementation biomethane producer and methane distributor).

Scenarios	As Is		To Be—65%SC		To Be—65%SC 40% OFMSW		To Be—110 kg/Inhabitant	
	Size	NPV	Size	NPV	Size	NPV	Size	NPV
Torino	500	10,473	500	13,122	500 [2]	20,946	500	17,488
			+250				+400	
Milano	500 [2] +300	24,837	500 [3]	31,419	500 [3]	31,419	500 [2] +400	27,961
Verona	250	2649	300	3891	350	5390	250	2649
Venezia	350	5390	400	7015	400	7015	250	2649
Padova	250	2649	300	3891	300	3891	200	1147
Trieste	0	0	0	0	250	2649	200	1147
Genova	0	0	350	5390	500 +200	11,620	500	10,473
Bologna	200	1147	300	3891	500	10,473	400	7015
Firenze	450	8679	500	10,473	500	10,473	400	7015
Roma	500 [4] +400	48,907	500 [7] +250	75,960	500 [8]	83,784	500 [5] +450	61,044
Napoli	500	10,473	500 [2] +200	22,093	500 [2] +250	23,595	500 [2]	20,946
Bari	0	0	300	3891	450	8679	350	5390
Palermo	0	0	500 +250	13,122	500 +350	15,863	500 +200	11,620
Messina	0	0	0	0	300	3891	250	2649
Catania	0	0	250	2649	500	10,473	300	3891
		Size (m ³	/h); NPV (the	ousand €); [1	n] = number	of plants		

Table 11. Definition of plant size and calculation of economic profit (thousand €)—alternative scenario (joint implementation biomethane producer and methane distributor).

5. Conclusions

This study began with some assumptions, and its final goal was to provide quantitative results. The literature review confirmed that biomethane is able to support the decarbonization of the transport sector, providing a CE model by which environmental and economic improvements are reached. While the role of cities is essential, a methodology was presented to policymakers based on the following steps: (i) the definition of the condition in which biomethane was profitable in the baseline scenario; (ii) the provision of alternative scenarios in which economic feasibility was demonstrated as a function of critical variables; (iii) the definition of the minimum biomethane plant size; (iv) the quantification of the biomethane potential as a function of future scenarios; and (v) the integration of the waste hierarchy into future choices, in order to support the integration of renewable energy management and waste management. Some results were obtained, but more analyses must be conducted. In particular, it would be extremely useful to analyze the specific context of a single municipality. At the same time, while Italy is an excellent case for research (as it is characterized by a large number of NGVs and a high amount of imported gas), this work proposed a model that could be applied in several contexts. Additionally, both the literature review and real experiences support the conclusion that Sweden represents the best-practice to follow. The definition of conditions in which biomethane is profitable represents a key element analyzed by investors. In addition, the literature also shows that the environmental impacts of organic waste recovered by the AD process that is upgraded in order to obtain biomethane represents a good option. The reduction of GHG emissions is extremely significant. Biomethane is a viable substitute for methane as a green fuel, and, in this way, NGVs can be fueled by a sustainable resource.

This work confirms that the recovery of biomethane from organic waste is profitable in plants at a minimum size of 200 m³/h. The new decree resolves some issues, and it defines the value of CICs. Subsidies account for approximately 40% of the discounted cash inflows, confirming their key role in determining the economic feasibility of plants. However, the same decree also has some issues. CICs

are recognized only for the first 10 years of biomethane activity, and, after this date, their price is not guaranteed. In addition, corrective coefficients proposed to support the use of some substrates (see double counting) and the premium to realize new points of methane distribution are not provided after 10 years. Thus, there is a risk that the revenues will exceed the costs for the first period of activity, and the opposite situation may occur in the second period because of the inflation of operative costs. This aspect must be monitored.

Policymakers have great interest in the transport sector, because there is a serious risk that the European target will not be reached. The development of advanced biofuels is a sustainable choice. The reduction of GHG emissions obtained by NGVs fueled by green gas corresponds to 12.6 tCO₂.eq for m³/h of biomethane capacity, annually. The profits derived from the application of biomethane plants are 408 K€ and 7942 K€ for 200 m³/h and 500 m³/h, respectively. DPBT is less than 1 year. Alternative scenarios are realized, considering variations in the CIC value after 10 years, the selling price of biomethane during the lifetime of the project, and the net revenues linked to the management of municipal waste. Also, a joint implementation model between a biomethane producer and methane distributor is proposed, with a significant increase of NPV equal to 1147 K€ and 10,473 K€ for 200 m³/h and 500 m³/h, respectively.

The final goal of this work was to define the potential economics of biomethane generated by managing the OFMSW in 15 Italian municipalities that, collectively, represent approximately 16.4% of the Italian population. Considering the future scenarios in which the separated rate collection will be fixed to 65%, the action plan obtained the following results: 80.4 to 102.8 million m³ of biomethane produced annually, 127 to 162 million kgCO₂.eq avoided annually, and an NPV of 135 to 187 million \notin over the lifetime of the projects. This latter value rises to 183 to 234 million \notin when the biomethane producer is also the methane distributor. Economic opportunities and environmental improvements are translated into clean energy, which is produced internally and applied to the transport sector. Each municipality has this potential, and both material and energy perspectives are valorized, providing a clear opportunity for increased sustainability.

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Nomenclature							
AD	anaerobic digestion	OFMSW	organic fraction municipal solid waste				
BIO-CNG	biomethane	PBT	payback time				
CE	circular economy	R ^{ofm}	optimistic scenario treatment of MSW				
CIC	certificates of emission of biofuel in consumption	R ^{ofm} _{pes}	pessimistic scenario treatment of MSW				
CNG	compressed natural gas	$\begin{array}{l} R_{opt}^{pos} \\ R_{pes}^{pos} \\ R_{opt}^{sel} \\ R_{pes}^{sel} \end{array}$	optimistic scenario point of sale				
CO ₂ eq	carbon dioxide equivalent	R ^{pos} _{pes}	pessimistic scenario point of sale				
DPBT	discounted payback time	R ^{sel} _{opt}	optimistic scenario selling price biomethane				
EU	European Union	R ^{sel} pes	pessimistic scenario selling price				
GHG	greenhouse gas		biomethane				
GHG _{BIOCNG}	emissions released by BIO-CNG	R ^{sub} opt R ^{sub} pes	optimistic scenario subsidies 11th year				
GHG _{CNG}	emissions released by CNG	R _{pes}	pessimistic scenario subsidies 11th year				
MS	member state	RES	renewable energy source				
MSW	municipal solid waste	RES-T	renewable energy source in the transport sector				
NGV	natural gas vehicle	RGHG _{NGV} .	reduction of GHG operated by NGVs				
NPV	net present value	%SC	separated collection rate				
NPV/Size	ratio between NPV and biomethane						
	plant size						

Nomenclature and Abbreviations

Nomenclature

Abbreviations

Abbreviations			
$1^{\circ}s$	biogas production	n _{debt}	period of loan
2°s	Upgrading	n _{oh}	number of operating hours
3°s	compression and distribution	n _{op}	number of operators
c _c	corrective coefficient (double counting)	n _s	period of subsidies (biomethane)
$C_{df}^{1^{\circ}s}$	depreciation fund (1°s)	Ot	discounted cash outflows
$\begin{array}{c} C_{df}^{1^{\circ}s} \\ C_{df,t}^{2^{\circ}s} \\ C_{e,t}^{2^{\circ}s} \\ C_{e,t}^{2^{\circ}s} \\ C_{e,t}^{2^{\circ}s} \\ C_{e,t}^{2^{\circ}s} \\ C_{e}^{2^{\circ}s} \\ C_{i,t}^{2^{\circ}s} \\ C_{i,t}^{1^{\circ}s} \\ C_{i,t}^{2^{\circ}s} \\ C_{i,t}^$	depreciation fund (2°s)	p_b^u	unitary potential of biomethane
$C_{a,t}^{1^{\circ}s}$	electricity cost (1°s)	P _{biomethane}	selling price of biomethane
$C^{2^{\circ}s}$	electricity cost (2°s)	p_{cic}^{u}	price of CIC (after 11 th year)
u,1°s	unitary electricity consumption (1°s)	r _{cic}	unitary price of CO_2
u,2°s	unitary electricity consumption (2°s)		price of compost
$c_{e}^{1^{\circ}s}$		-	
$C_i^{2^\circ s}$	insurance cost (1°s)	P _{df}	percentage of depreciation fund
C- o i,t	insurance cost $(2^{\circ}s)$	pe	unitary price of electricity
C_{inv}^{uig}	investment cost (digestate)	p_i	percentage of insurance cost
C ^{dis} _{inv}	investment cost (distribution)	$p_{mo}^{1^{\circ}s}$	percentage of maintenance and overhead cost (1°s)
Cl	labor cost	p _{mo} ^{2°s}	percentage of maintenance and overhead cost (2°s)
$C_l^{u,a}$	unitary labor cost	P _{PTV}	price in the virtual trading point
C_{lcs}	loan capital share cost	P _{tax}	percentage of taxes
Clie	loan interest share cost	p_z^u	unitary price of zeolite
$C^{u,1^{\circ}s}$	unitary investment cost $(1^{\circ}s)$	PTV	virtual trading point
$\begin{array}{c} C_{iny}^{u,1^{\circ}s} \\ C_{inv}^{u,2^{\circ}s} \\ C_{inv}^{1^{\circ}s} \\ C_{mo}^{1^{\circ}s} \\ C_{mo}^{2^{\circ}s} \end{array}$	unitary investment cost (2°s)	R	opportunity cost
$C^{1^{\circ}s}$	maintenance and overhead cost (1°s)		recovery rate (biomethane)
$C^{2^{\circ}s}$	maintenance and overhead cost (1°)		recovery rate (CO_2)
	consumption of NGV		revenues by compost
c _{NGV} C _o ^{dig}	_	R _{compost}	interest rate on loan
C _o	operative cost (digestate)	r_d $R_t^{CO_2}$	
Codis	operative cost (distribution)	R _t ^{OFMSW}	revenues by selling CO ₂
C _s	substrate cost	Rt DFMSW	revenues by treatment of the OFMSW
C ^u _s	unitary substrate cost	R ^{OFMSW} gross,t	gross revenues by the OFMSW
C _t OFMSW	cost of the OFMSW	R _{gross,t} R _t	revenues by selling biomethane
C _{tax}	taxes cost	R_t^{t}	revenues by subsidies
Cts	transport cost of substrates	Q _{biogas}	quantity of biogas
C ^u _{ts}	unitary transport cost of Substrate	Q _{biogas}	nominal quantity of biogas
C _{gv}	generic variable cost	Qbiomethane	
cf _{CO2}	conversion factor (CO_2)	Q _{CO₂}	quantity of CO_2
cf _{gas}	conversion factor (gas)		
CF	cash flow	$\begin{array}{c} Q_{compost} \\ Q_{biomethane}^{nom} \end{array}$	nominal quantity of biomethane
C1	certificates of emission of biofuel in	≪biomethane	
CIC	consumption	Q _{OFMSW}	quantity of the OFMSW
Dig	Digestate	Qz	quantity of zeolite
DPBT	discounted payback time	S _{biogas}	plant size (biogas)
It	discounted cash inflows	S _{biomethane}	plant size (biomethane)
i ^u cic	unitary subsidy (biomethane)	t	time of the cash flow
Inf	rate of inflation	%CH ₄	percentage of methane
l _{bs}	losses in the biogas system	%CO ₂	percentage of carbon dioxide
l _{us}	losses in the upgrading system	%rp _{PTV}	percentage reduction of the price in the virtual trading point
n	lifetime of investment		01

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